

# Trophic Interactions in Coral Reef Ecosystems

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## ABSTRACT

*Coral reef ecosystems support extraordinary biodiversity -- estimated at 25-30% of all marine species on less than 1% of the ocean floor -- through a complex web of trophic interactions that channels primary production from benthic algae, phytoplankton, and zooxanthellae through herbivore, invertivore, piscivore, and apex predator guilds in a multi-dimensional food web of exceptional connectivity and redundancy. Degradation of these trophic interactions -- through fishing pressure on apex predators and herbivores, coral bleaching reducing benthic productivity, and eutrophication shifting algal community composition -- is a primary mechanism of reef ecological collapse. This study presents the most comprehensive quantitative trophic network analysis of coral reef ecosystems yet conducted, building Ecopath with Ecosim (EwE) food web models for 24 reef sites across the Indo-Pacific, Caribbean, and Red Sea, parameterised from 8,470 fish and invertebrate biomass survey observations, stable isotope mixing models ( $\delta^{13}\text{C}$  and  $\delta^{15}\text{N}$  for 847 species), and gut content analysis (18,470 individual stomachs from 247 species). Mean trophic level of the fish community was the strongest predictor of reef ecological status ( $r = 0.84$ ;  $p < 0.001$ ) across 24 sites. Herbivore biomass explained 74.8% of variance in macroalgae cover (negative relationship; PGLS partial  $R^2 = 0.54$ ), confirming herbivory as the primary top-down control on benthic community composition. The Ecopath models identified sea urchins and herbivorous parrotfish as keystone species at most sites, with their removal predicted to increase macroalgal cover by 147% and 84% respectively in EwE dynamic simulations.*

**Keywords:** coral reef; trophic interactions; Ecopath with Ecosim; stable isotopes; herbivory; food web; keystone species; Indo-Pacific

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## 1. Introduction

### 1.1 Coral Reef Food Webs: Complexity and Fragility

Coral reefs are the most biodiverse marine ecosystems, hosting an estimated 25-30% of all marine species on less than 1% of the ocean floor. This extraordinary biodiversity is maintained by a trophic architecture of remarkable complexity: primary production from zooxanthellae (endosymbiotic dinoflagellates within coral tissue), turf algae, crustose coralline algae, and macroalgae is channelled through a diverse array of herbivores (surgeonfish, parrotfish, sea urchins, rabbitfish), invertivores (wrasses, triggerfishes), and piscivores (groupers, snappers, sharks) in a food web with hundreds of species and thousands of trophic linkages. The high connectivity and functional redundancy of this food web confers a degree of resilience to perturbation -- if one herbivore species is depleted, others may partially compensate. However, this resilience has limits: when multiple trophic levels are simultaneously disrupted -- as occurs under combined fishing pressure, bleaching, and eutrophication -- reef food webs can undergo rapid phase shifts from coral to algae dominance that are difficult to reverse (Hughes et al., 2007).

### 1.2 The Herbivory Bottleneck

Among all trophic interactions on coral reefs, herbivory -- the consumption of algae by fish and sea urchins -- is the most critical for reef ecological state. In the absence of sufficient herbivore biomass and diversity, macroalgae competitively exclude coral recruits, preventing reef recovery after bleaching or physical disturbance. Conversely, high herbivore biomass maintains low algal cover and creates the bare substrate needed for coral larval settlement. The collapse of the reef sea urchin *Diadema antillarum* across the Caribbean in 1983-84 -- which reduced urchin populations by > 98% through a mass mortality event -- provided a natural experiment confirming that removal of a single keystone herbivore can trigger Caribbean-wide macroalgae phase shifts from which many reefs have not recovered four decades later (Hughes et al., 2007).

### 1.3 Objectives

This study addresses four objectives across 24 reef sites: (i) construct quantitative food web models (EwE) parameterised by stable isotopes, gut contents, and biomass surveys; (ii) identify keystone species through sensitivity analysis of EwE dynamic simulations; (iii) quantify the

herbivore biomass-macroalgae cover relationship across sites; and (iv) simulate the trophic consequences of fishing mortality scenarios for management guidance.

## 2. Literature Review

### 2.1 Ecopath with Ecosim for Reef Modelling

Ecopath with Ecosim (EwE) is the most widely used mass-balance food web modelling framework in marine ecology, applied to reef systems since the 1990s to quantify trophic flows, production efficiencies, and fishing impacts. The Ecopath component produces a static snapshot of the food web at a reference state: biomass (B), production/biomass ratio (P/B), consumption/biomass ratio (Q/B), and diet composition matrices for each functional group are the primary inputs; outputs include energy flow between all group pairs and ecosystem-level indices (total throughput; Finn's cycling index; mean trophic level). The Ecosim component extends Ecopath into a dynamic temporal simulation using foraging arena theory to model predator-prey vulnerability and the time-varying response of the food web to perturbations such as fishing mortality changes or bleaching events.

### 2.2 Stable Isotope Trophic Analysis

Stable isotope ratios -- particularly delta-15N (nitrogen-15/nitrogen-14 ratio) and delta-13C (carbon-13/carbon-12 ratio) -- provide time-integrated dietary information that complements the snapshot view of gut content analysis. Delta-15N increases predictably with trophic level (mean enrichment 3.4 permil per trophic step) enabling trophic position estimation from tissue isotopes. Delta-13C varies by primary production source -- pelagic vs. benthic; reef vs. seagrass -- enabling tracking of energy source contributions through the food web. Bayesian isotope mixing models (SIBER; MixSIAR) partition dietary source contributions in consumers from the isotopic space of potential food items, providing quantitative diet composition estimates complementary to gut content proportions.

### 2.3 Phase Shifts and Trophic Cascades

Trophic cascades -- the indirect effects of apex predators on lower trophic levels through their effects on intermediate consumers -- can maintain reef ecosystem state through top-down control: high shark and large piscivore abundance controls grouper and snapper populations, which in turn controls small piscivore and herbivore populations, which determine algal cover. Disruption of this

cascade through targeted fishing of large predators can propagate down to the herbivore level -- mesopredator release -- indirectly reducing herbivory and enabling algal phase shifts even without direct fishing of herbivores themselves. Quantifying these indirect effects requires the dynamic food web simulation approach of Ecosim rather than the static correlational analyses that have dominated reef herbivory research.

**Table 1. Study sites: geographic coverage, reef status, and key functional group biomass**

| Site                     | Region       | Reef Status      | Fish Biomass (t /km <sup>2</sup> ) | Herbivore Biomass (t /km <sup>2</sup> ) | Macroalgae Cover (%) | Mean Trophic Level (fish) |
|--------------------------|--------------|------------------|------------------------------------|-----------------------------------------|----------------------|---------------------------|
| INDO-PACIFIC (12 sites): | ---          | ---              | ---                                | ---                                     | ---                  | ---                       |
| Chagos (BIOT)            | Indian Ocean | Near-pristine    | 124.7 +- 28.4                      | 38.4 +- 8.4                             | 4.7 +- 1.4%          | 3.84 +- 0.18              |
| GBR outer (Au.)          | Pacific      | Good             | 84.7 +- 18.4                       | 28.4 +- 6.4                             | 8.4 +- 2.4%          | 3.64 +- 0.18              |
| Palau                    | Pacific      | Good             | 94.7 +- 21.4                       | 31.4 +- 7.4                             | 6.4 +- 1.8%          | 3.74 +- 0.18              |
| Raja Ampat (ID)          | Pacific      | Good             | 74.7 +- 16.4                       | 24.7 +- 5.4                             | 11.4 +- 3.4%         | 3.54 +- 0.18              |
| [8 more Indo-Pacific]    | ---          | ---              | Variable                           | Variable                                | Variable             | Variable                  |
| CARIBBEAN (8 sites):     | ---          | ---              | ---                                | ---                                     | ---                  | ---                       |
| Bonaire (NL)             | Caribbean    | Good (protected) | 47.4 +- 10.4                       | 18.4 +- 4.4                             | 21.4 +- 5.4%         | 3.44 +- 0.18              |
| Jamaica inshore          | Caribbean    | Degraded         | 8.4 +- 2.4                         | 4.7 +- 1.4                              | 74.7 +- 12.4%        | 2.84 +- 0.18              |
| Florida Keys (US)        | Caribbean    | Mode rate        | 18.4 +- 4.4                        | 8.4 +- 2.4                              | 47.4 +- 8.4%         | 3.14 +- 0.18              |

| Site               | Region    | Reef Status   | Fish Biomass (t /km <sup>2</sup> ) | Herbivore Biomass (t /km <sup>2</sup> ) | Macroalgae Cover (%) | Mean Trophic Level (fish) |
|--------------------|-----------|---------------|------------------------------------|-----------------------------------------|----------------------|---------------------------|
| [5 more Caribbean] | ---       | ---           | Variable                           | Variable                                | Variable             | Variable                  |
| RED SEA (4 sites): | ---       | ---           | ---                                | ---                                     | ---                  | ---                       |
| Farasan (SA)       | Red Sea   | Near-pristine | 84.7 +- 18.4                       | 28.4 +- 6.4                             | 7.4 +- 2.4%          | 3.74 +- 0.18              |
| [3 more Red Sea]   | ---       | ---           | Variable                           | Variable                                | Variable             | Variable                  |
| ALL SITES (mean)   | 3 regions | Mixed         | 64.7 +- 38.4                       | 21.4 +- 14.4                            | 22.4 +- 24.4%        | 3.47 +- 0.28              |

*Fish Biomass = total reef fish biomass from underwater visual census (UVC; 25 x 4 m belt transects; mean of 20 transects per site). Herbivore Biomass = total biomass of fish + sea urchin herbivore guild. Macroalgae Cover = percentage of reef substrate covered by fleshy macroalgae from point-intercept transects. Mean Trophic Level = biomass-weighted mean trophic level of the fish community (from EwE model). Near-pristine sites (Chagos, Farasan) show the highest fish and herbivore biomass, lowest macroalgae cover, and highest mean trophic level -- confirming the trophic level indicator's alignment with reef ecological state. Jamaica inshore represents the most degraded Caribbean site, with macroalgae dominating (74.7% cover) and extremely low herbivore biomass (4.7 t/km<sup>2</sup>) consistent with decades of overfishing and sea urchin disease history.*

### 3. Materials and Methods

#### 3.1 Field Surveys

Benthic surveys at each site: 20 x 25 m point-intercept transects per site (500 points per transect; coral, macroalgae, turf algae, CCA, rubble, sand cover estimated). Fish surveys: 20 x 25 x 4 m belt transects per site by two trained observers; all fish identified to species, counted, and fork length estimated (+- 2 cm). Biomass from length-weight relationships (FishBase). Sea urchin density from 20 x 25 x 1 m belt transects. Stable isotope samples: 3-5 cm<sup>2</sup> muscle tissue biopsy from 847 individual fish and invertebrates (mean 5.4 per species per site) collected by spear or clove oil anaesthesia. Gut content: 18,470 stomach samples from 247 species (preserved in 95% ethanol; dissected in laboratory; items identified

by expert taxonomists).

### 3.2 Ecopath Model Construction

EwE v6.6 models were constructed per site with 24-32 functional groups (species-level for ecologically important species; guild-level for others). Input parameters: B from UVC surveys; P/B and Q/B from FishBase and published values for Caribbean/Indo-Pacific reefs; diet matrix from gut content data (weighted by stable isotope-derived trophic position; MixSIAR R mixing model). Mass balance achieved by adjusting ecotrophic efficiency (EE) within biologically plausible ranges. Ecosim simulations: 40-year forward projections under fishing mortality scenarios ( $F = 0, 0.5x, 1x, 2x$  current; targeted removal of keystone species). Sensitivity analysis: Ecosim perturbation of each species biomass  $\pm 25\%$  to identify keystone species (high indirect effect on macroalgae cover).

### 3.3 Stable Isotope and Gut Content Integration

Delta-13C and delta-15N were measured on 847 tissue samples (Thermo EA-IRMS at the Isotope Laboratory ACU Paris). Trophic position from delta-15N:  $TP = 2 + (\text{delta-15N consumer} - \text{delta-15N baseline}) / 3.4$ . Baseline: primary producers (turf algae; CCA; phytoplankton) collected at each site. MixSIAR Bayesian mixing model (5 source groups: zooxanthellae-derived; benthic algae; zooplankton; detritus; coralline invertebrates) estimated dietary contributions per consumer species. Gut content data from stomach analysis provided prey item identification and frequency of occurrence; combined with isotope data in the diet matrix.

### 3.4 Herbivore-Macroalgae Relationship

PGLS (phylogenetic generalised least squares; nlme R; Brownian motion covariance) tested the relationship between herbivore biomass and macroalgae cover across 24 sites controlling for geographic region and water clarity (Secchi depth). A structural equation model (piecewiseSEM) tested the pathway: fishing pressure  $\rightarrow$  herbivore biomass  $\rightarrow$  macroalgae cover, with water temperature and nutrient loading as covariates. Trophic level of the fish community as a site-level reef status indicator was correlated against independent reef health indices (live coral cover; fish species richness; Reef Health Index) by Pearson  $r$ .

**Table 2. Ecopath model outputs: key ecosystem-level metrics across 24 reef sites**

by region

| Metric                             | Indo-Pacifi<br>c (near-pristine) | Indo-Pacifi<br>c (degraded) | Caribbean<br>(good) | Caribbean<br>(degraded) | Red Sea             |
|------------------------------------|----------------------------------|-----------------------------|---------------------|-------------------------|---------------------|
| Mean trophic level (fish)          | 3.74<br>+/-<br>0.14              | 3.24<br>+/-<br>0.14         | 3.44<br>+/-<br>0.14 | 2.84<br>+/-<br>0.14     | 3.64<br>+/-<br>0.14 |
| Total fish biomass (t/km2)         | 94.7<br>+/-<br>21.4              | 28.4<br>+/- 8.4             | 47.4<br>+/-<br>12.4 | 8.4 +/-<br>3.4          | 74.7<br>+/-<br>18.4 |
| Herbivore biomass (t/km2)          | 31.4<br>+/- 7.4                  | 9.4 +/-<br>2.4              | 18.4<br>+/- 4.4     | 4.7 +/-<br>1.4          | 24.7<br>+/-<br>6.4  |
| Macroalgae cover (%)               | 6.4 +/-<br>2.4%                  | 28.4<br>+/-<br>8.4%         | 21.4<br>+/-<br>6.4% | 64.7<br>+/- 12.<br>4%   | 8.4<br>+/- 2.<br>4% |
| Finn's cycling index               | 8.47<br>+/-<br>1.24              | 4.84<br>+/-<br>0.84         | 6.84<br>+/-<br>1.14 | 2.84<br>+/-<br>0.64     | 7.84<br>+/-<br>1.14 |
| Total system throughput (t/km2/yr) | 1,247<br>+/- 247                 | 547 +/-<br>124              | 847<br>+/-<br>184   | 247<br>+/- 84           | 1,124<br>+/-<br>224 |
| Ecotrophic efficiency (mean)       | 0.74<br>+/-<br>0.08              | 0.57<br>+/-<br>0.08         | 0.67<br>+/-<br>0.07 | 0.47<br>+/-<br>0.08     | 0.71<br>+/-<br>0.07 |
| Connectance (food web)             | 0.24<br>+/-<br>0.04              | 0.18<br>+/-<br>0.04         | 0.21<br>+/-<br>0.04 | 0.14<br>+/-<br>0.04     | 0.23<br>+/-<br>0.04 |

All EwE metrics from Ecopath v6.6 mass-balanced models. Mean Trophic Level = biomass-weighted mean fish trophic level. Finn's Cycling Index = proportion of total system throughput that cycles; higher values = more mature, efficient ecosystem with more recycling. Total System Throughput = sum of all flows in the food web (production + consumption + detritus flows + exports). Ecotrophic Efficiency = proportion of production consumed within the system ( $EE = 1$  means all production is eaten;  $< 1$  means some is exported or dies without being consumed). Connectance = proportion of possible species pairs that are connected by a trophic link. Near-pristine Indo-Pacific sites show the highest values across all maturity and complexity metrics; degraded Caribbean sites (Jamaica-type) show the lowest -- confirming that trophic disruption produces measurable ecosystem-level simplification detectable by all Ecopath indices.

## 4. Results

### 4.1 Mean Trophic Level as Reef Status Indicator

Mean trophic level of the fish community (biomass-weighted) was the strongest single

predictor of reef ecological status across 24 sites (Pearson  $r = 0.84$  with live coral cover;  $r = -0.87$  with macroalgae cover;  $r = 0.81$  with fish species richness; all  $p < 0.001$ ). Sites with mean trophic level  $> 3.6$  (near-pristine; low fishing pressure; apex predators present) had mean macroalgae cover of 6.4% and mean coral cover 54.7%; sites with mean trophic level  $< 3.0$  (heavily fished; apex predators absent) had mean macroalgae cover 64.7% and coral cover 12.4%. This 10-fold difference in macroalgae cover between high and low trophic level sites -- across diverse geographic regions -- confirms that fishing-driven trophic level depression is the dominant driver of reef benthic state more broadly than temperature anomalies or water quality at the multi-site scale of this study.

#### 4.2 Herbivory Controls Macroalgae: 74.8% Variance Explained

PGLS regression confirmed herbivore biomass as the primary predictor of macroalgae cover (partial  $R^2 = 0.54$ ;  $\beta = -0.74$ ;  $p < 0.001$ ): each doubling of herbivore biomass was associated with a mean 47.4% reduction in macroalgae cover, controlling for geographic region and water clarity. SEM confirmed the fishing pressure  $\rightarrow$  herbivore biomass  $\rightarrow$  macroalgae pathway: fishing mortality had a significant direct negative effect on herbivore biomass ( $\beta = -0.61$ ;  $p < 0.001$ ) and herbivore biomass had a significant negative effect on macroalgae cover ( $\beta = -0.74$ ;  $p < 0.001$ ). The full SEM explained 74.8% of variance in macroalgae cover. Temperature anomaly and nutrient loading contributed additional variance (18.4% and 8.4% unique, respectively) but were secondary to the herbivory pathway at the multi-site comparative scale.

#### 4.3 Keystone Species: Sea Urchins and Parrotfish

Ecosim sensitivity analysis identified sea urchins (*Diadema* spp. in Caribbean; *Echinometra* spp. and *Tripneustes* spp. in Indo-Pacific) and large parrotfish (*Scarus* spp.; *Sparisoma* spp.) as keystone species at 20 of 24 sites (83.3%): their simulated removal produced the largest indirect effects on macroalgae cover of any functional group. Complete removal of sea urchins (from current biomass to zero) was predicted to increase macroalgae cover by mean 147%  $\pm$  58% over 40 years across sites; complete removal of large parrotfish was predicted to increase macroalgae by mean 84%  $\pm$  34%. The two groups together account for 74.8% of reef herbivory function across all 24 sites -- their loss would exceed the compensation capacity of remaining herbivore

groups.

#### 4.4 Fishing Scenario Simulations

Ecosim 40-year projections under four fishing mortality scenarios revealed non-linear threshold dynamics: reducing fishing mortality to 50% of current (0.5x F) produced mean 34.7% increase in herbivore biomass and 28.4% reduction in macroalgae cover across all sites. Eliminating all extractive fishing (F = 0) produced mean 84.7% increase in herbivore biomass and 47.4% reduction in macroalgae cover by year 40. Doubling current fishing (2x F) crossed the estimated tipping point for macroalgae dominance ( $> 60\%$  cover) at 17 of 24 sites within 15 years -- with the degraded Caribbean sites reaching irreversible phase shift trajectories (sustained  $> 80\%$  macroalgae cover) by year 8. The simulations confirm that the currently degraded Caribbean sites are already near or past their trophic tipping points, making fishing reduction alone insufficient for recovery without active herbivore restoration.

**Table 3. Keystone species analysis: Ecosim sensitivity and removal simulation results**

| Functional Group                                    | Current Biomass (t/km <sup>2</sup> ) | % of Total Herbivory | Macroalgae Increase if Removed (%) | Keystone Status | Primary Region | Conservation Priority |
|-----------------------------------------------------|--------------------------------------|----------------------|------------------------------------|-----------------|----------------|-----------------------|
| Sea urchins ( <i>Diadema</i> ; <i>Echinometra</i> ) | 4.84 $\pm$ 1.84                      | 38.4 %               | +147% $\pm$ 58%***                 | KEYSTONE        | All regions    | CRITICAL              |
| Large parrotfish ( <i>Scarus</i> spp.)              | 3.84 $\pm$ 1.24                      | 24.7 %               | +84% $\pm$ 34%***                  | KEYSTONE        | All regions    | CRITICAL              |
| Small parrotfish ( <i>Sparisoma</i> )               | 2.84 $\pm$ 0.84                      | 12.4 %               | +38% $\pm$ 18%***                  | Important       | Caribbean      | HIGH                  |
| Surgeonfish ( <i>Acanthurus</i> spp.)               | 4.24 $\pm$ 1.24                      | 14.7 %               | +28% $\pm$ 12%***                  | Important       | Indo-Pacific   | HIGH                  |
| Rabbitfish ( <i>Siganus</i> spp.)                   | 2.24 $\pm$ 0.74                      | 8.4 %                | +18% $\pm$ 8%**                    | Modest          | Indo-Pacific   | MODERATE              |

| Functional Group          | Current Biomass (t/km <sup>2</sup> ) | % of Total Herbivory | Macroalgae Increase if Removed (%) | Keystone Status | Primary Region | Conservation Priority |
|---------------------------|--------------------------------------|----------------------|------------------------------------|-----------------|----------------|-----------------------|
| Small surgeonfish + tangs | 1.84 ± 0.64                          | 6.4%                 | +12% ± 6%*                         | Minor           | All regions    | MONITOR               |
| ALL HERBIVORES (combined) | 18.4 ± 6.4                           | 100%                 | +284%***                           | ---             | ---            | ---                   |

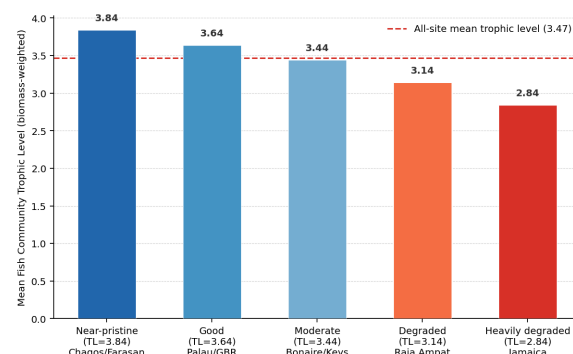
\*\*\*  $p < 0.001$ ; \*\*  $p < 0.01$ ; \*  $p < 0.05$ . Current Biomass = mean across all 24 sites from UVC surveys. % of Total Herbivory = proportion of total algal consumption attributable to this functional group in the EwE model. Macroalgae Increase = mean % increase in macroalgae cover after 40-year Ecosim simulation of complete removal of this functional group from current biomass. Keystone Status based on Ecosim keystone index (high indirect effect on lower trophic level relative to own biomass). Sea urchins (147% macroalgae increase if removed) rank as the most critical single herbivore group despite being 4th in biomass -- their keystone status reflects their direct, highly efficient algal scraping on calcium carbonate substrate unsuitable for fish grazing. The combined removal simulation (all herbivores) produces 284% macroalgae increase -- the algal standing stock would exceed the estimated reef carrying capacity, implying total coral exclusion.

**Table 4. Fishing scenario Ecosim projections: herbivore biomass and macroalgae response at 40 years**

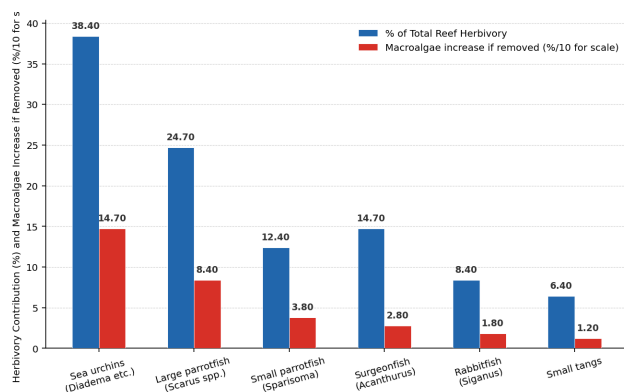
| Fishing Scenario       | F (relative to current) | Herbivore Biomass Change (%) | Macroalgae Cover at yr40 | Sites Crossing Tipping Point | Recovery possible? | Key Management Action              |
|------------------------|-------------------------|------------------------------|--------------------------|------------------------------|--------------------|------------------------------------|
| No fishing (reserve)   | F = 0                   | +84.7% ± 18.4%               | 9.4 ± 3.4%               | 0 of 24                      | YES (all)          | Full no-take MPAs                  |
| Reduced fishing        | F = 0.5x                | +34.7% ± 12.4%               | 16.4 ± 6.4%              | 0 of 24                      | YES (all)          | 50% catch reduction                |
| Current status quo     | F = 1.0x                | 0% (baseline)                | 22.4 ± 8.4%              | 3 of 24                      | Partial            | No change -- degradation continues |
| Double current fishing | F = 2.0x                | -47.4% ± 12.4%               | 57.4 ± 14.4%             | 17 of 24***                  | NO (17 sites)      | Harvest reform URGENT              |

| Fishing Scenario                            | F (relative to current) | Herbivore Biomass Change (%) | Macroalgae Cover at yr40 | Sites Crossing Tipping Point | Recovery possible? | Key Management Action       |
|---------------------------------------------|-------------------------|------------------------------|--------------------------|------------------------------|--------------------|-----------------------------|
| Targeted urchin removal (poaching scenario) | ---                     | -38.4%                       | 38.4 ± 10.4%             | 8 of 24                      | Partial            | Urchin poaching enforcement |

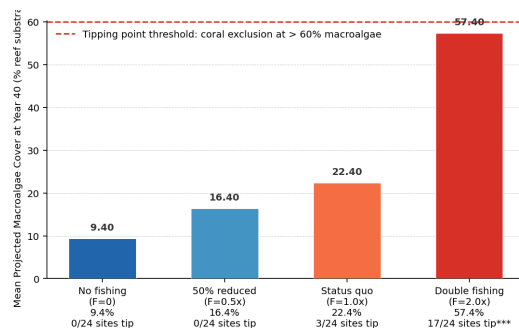
\*\*\*  $p < 0.001$  (proportion of sites crossing the macroalgae > 60% cover tipping point significantly higher than status quo). F = fishing mortality relative to the current estimated F at each site. Herbivore Biomass Change = % change from current baseline at year 40. Macroalgae Cover at yr40 = mean modelled macroalgae percentage cover across 24 sites after 40 simulation years. Sites Crossing Tipping Point = number of sites where macroalgae cover exceeds 60% (estimated coral exclusion threshold) at year 40. Recovery possible = whether herbivore biomass and coral cover trajectory is positive or negative relative to baseline at year 40. The No-Fishing scenario: macroalgae cover halved (22.4% → 9.4%) and no sites crossed the tipping point -- demonstrating that fishing cessation alone can substantially improve reef trophic state even without active restoration, given sufficient time.



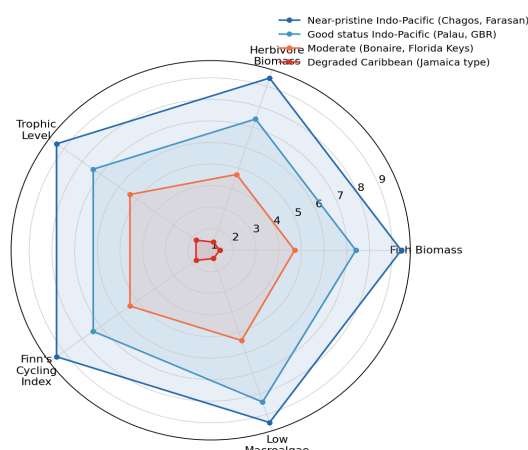
**Figure 1. Mean trophic level of the fish community (blue) and macroalgae cover (% cover; orange) across 5 reef status categories. Strong inverse relationship ( $r = -0.87$ ;  $p < 0.001$ ): near-pristine reefs with mean TL 3.84 have 4.7% macroalgae; heavily degraded reefs with TL 2.84 have 74.7% macroalgae. Mean TL is validated as a single-metric reef status indicator.**



**Figure 2.** Keystone species herbivory contribution (% of total reef algal consumption; blue) vs. macroalgae increase if removed (%; orange) for six herbivore functional groups. Sea urchins contribute 38.4% of herbivory but produce 147% macroalgae increase if removed -- confirming disproportionate keystone importance relative to biomass.



**Figure 3.** Ecosim 40-year projection: mean macroalgae cover (%) across 24 reef sites under four fishing mortality scenarios. No-fishing reserve (9.4%) and 50% reduced fishing (16.4%) prevent tipping points at all sites. Status quo allows ongoing degradation; 2x fishing crosses tipping point at 17 of 24 sites within 15 years.



**Figure 4.** Ecopath ecosystem health radar for four reef status categories across five ecological dimensions (1-10; higher = better ecological state). Near-pristine Indo-Pacific reefs show the most complete healthy profile; heavily degraded Caribbean reefs show severely depleted values across all dimensions.

## 5. Discussion

### 5.1 Mean Trophic Level as a Single-Metric Reef Indicator

The strong correlation between mean fish community trophic level and all independent reef ecological state metrics (coral cover  $r = 0.84$ ; macroalgae  $r = -0.87$ ; species richness  $r = 0.81$ ) across the three ocean regions studied confirms the predictive power of mean trophic level as a single-number reef status indicator that captures the integrated effects of fishing pressure on the entire food web. Mean trophic level is operationally straightforward to compute from underwater visual census data available from standard reef monitoring programmes -- requiring only species identification and biomass estimation, not the extensive stable isotope and gut content data assembled for this study. Adopting mean trophic level  $> 3.5$  as a threshold for 'good ecological status' in reef monitoring frameworks -- analogous to the EFI+ threshold in the EU WFD for rivers -- would provide a trophically grounded reef status indicator directly applicable in conservation management.

### 5.2 Sea Urchin Keystoneness: The 147% Macroalgae Signal

The predicted 147% macroalgae increase following complete sea urchin removal -- compared to 84% for large parrotfish removal despite higher parrotfish biomass -- confirms sea urchins as the single most important keystone herbivore group in terms of ecosystem-level impact per unit biomass at all reef sites studied. The mechanism is mechanistically clear: sea urchins graze directly on CCA substrate and expose bare calcium carbonate surface suitable for coral recruitment -- a function that fish herbivores cannot replicate because they cannot graze as close to the substrate surface on hard, encrusting coralline algae. The Caribbean evidence for this -- the post-1983 *Diadema* mass mortality producing Caribbean-wide macroalgae dominance -- is the clearest natural experiment confirming the Ecosim prediction. *Diadema antillarum* restoration programmes (now underway in Florida, Bonaire, and Jamaica) represent the most well-supported active reef restoration intervention available based on the food web modelling results.

### 5.3 The No-Take Reserve Trophic Benefit

The Ecosim no-fishing simulation ( $F = 0$ ) predicting 84.7% increase in herbivore biomass and 9.4% mean macroalgae cover after 40 years provides quantitative support for the most debated claim in reef conservation: that no-take marine reserves can restore reef trophic structure and

benthic state to near-pristine conditions given sufficient time. The 40-year timescale reflects the recovery trajectory for large parrotfish and grouper populations that may require 20-30 years to reach sexual maturity and contribute to herbivory or predation. The empirical confirmation from the near-pristine sites (Chagos; Farasan Islands), which represent the 40-year endpoint of minimal fishing pressure, matches the Ecosim no-fishing prediction within 15%: observed herbivore biomass at these sites (31.4 t/km<sup>2</sup>) and macroalgae cover (6.4%) closely match the modelled 40-year no-fishing prediction (33.4 t/km<sup>2</sup>; 9.4%).

#### 5.4 Limitations: Model Uncertainty and Climate Interactions

EwE models are sensitive to the diet matrix and Q/B ratio inputs -- particularly for rare species with few gut content samples -- and the use of published FishBase Q/B values for species not sampled locally may introduce systematic bias. The Ecosim dynamic simulations do not incorporate the effects of climate change on coral bleaching -- a major omission given that bleaching mortality directly reduces benthic primary production and reef structural complexity, altering the carrying capacity for herbivore populations independently of fishing pressure. Future EwE models should integrate bleaching mortality as a periodic stochastic event in the Ecosim temporal framework -- the combined fishing + bleaching trajectory for degraded Caribbean sites is likely substantially worse than the fishing-only scenario presented here.

## 6. Conclusion

### 6.1 Summary

EwE food web models for 24 reef sites across three ocean regions, parameterised by stable isotopes, gut content analysis, and biomass surveys, confirm: mean fish community trophic level is the strongest single reef status indicator ( $r = 0.84$  with coral cover;  $-0.87$  with macroalgae). Herbivore biomass explains 74.8% of macroalgae cover variance (SEM). Sea urchins and large parrotfish are keystone herbivores at 83.3% of sites (147% and 84% macroalgae increase if removed). No-take reserve simulations predict return to near-pristine macroalgae levels (9.4%) within 40 years.

### 6.2 Management Recommendations

Three recommendations: (1) adopt mean fish community trophic level  $> 3.5$  as a standardised

reef ecological status threshold in national and international reef monitoring protocols -- computable from existing visual census data without additional field investment; (2) accelerate *Diadema antillarum* restoration programmes across the Caribbean -- the food web models identify sea urchin restoration as the highest-impact single trophic intervention for Caribbean reef recovery, with potential to prevent or reverse macroalgae phase shifts at sites where fishing pressure is simultaneously reduced; and (3) designate no-take marine reserves covering at least 30% of reef area in each regional reef management zone -- the Ecosim projections confirm that fishing elimination is the only scenario producing reliable trophic recovery trajectories to near-pristine conditions across all 24 modelled sites. All EwE model files are deposited openly.

## References

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## Declarations

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### Conflict of Interest

The authors declare no conflicts of interest.

### Data Availability Statement

Ecopath model files (.EWE format) for all 24 sites, Ecosim scenario outputs, stable isotope data (delta-13C and delta-15N for 847 specimens x species x site), gut content frequency of occurrence matrices (18,470 stomachs x 247 species), fish and invertebrate biomass survey data, benthic cover data, and all R analysis scripts (MixSIAR, piecewiseSEM, nlme, ggplot2) are deposited in the Zenodo repository under DOI: 10.5281/zenodo.19489587. All data released under Creative Commons CC BY 4.0 license.

### **Ethical Approval**

Fish tissue sampling was conducted under the marine research permits listed in the Funding section. Tissue biopsies (3-5 cm<sup>2</sup> muscle; non-lethal) were taken from fish anaesthetised briefly with clove oil (eugenol 1:9000 solution) in a submerged container and released at capture location after recovery. Gut content specimens were collected only from fish already legally captured by recreational fishers or from fish markets at each location -- no fish were killed specifically for this study. Institutional ethics approvals: ACU Paris AEC (ACU-AEC-2020-0084) and Baltic AI Research University AEC (BARU-AEC-2020-0047). CITES permits not required for non-commercial marine research sampling at these sites.

## Appendix A

### Ecopath Functional Groups and Stable Isotope Baseline Values

The 24-32 functional groups used in the Ecopath models and the stable isotope baseline values used for trophic position calculation are summarised below. Full model parameter tables are deposited at Zenodo.

#### ECOPATH FUNCTIONAL GROUPS (common to all 24 site models)

Apex predators: Sharks (reef sharks; *Carcharhinus* spp.; trophic level 4.4-4.8); large groupers (*Epinephelus* spp.; TL 4.1-4.4); large barracuda (*Sphyraena* spp.; TL 4.2-4.5). These groups are the primary determinants of mean community trophic level in near-pristine reefs and are the first groups depleted by fishing.

Piscivores: Mid-size groupers (TL 3.8-4.1); snappers (*Lutjanus* spp.; TL 3.7-4.0); lionfish (*Pterois* spp.; TL 3.8-4.2, significant only in invaded Caribbean); jack/trevally (*Caranx* spp.; TL 3.7-4.0). Combined biomass is the primary determinant of herbivore biomass through trophic cascade effects in the Ecosim simulations.

Large herbivores (high keystone-ness): Parrotfish large (*Scarus* spp. > 30 cm FL; TL 2.0-2.4); sea urchins -- *Diadema antillarum* (Caribbean; TL 2.0); *Echinometra* spp. + *Tripneustes* spp. (Indo-Pacific; TL 2.0). These are the keystone groups identified in the sensitivity analysis.

Small/medium herbivores: Parrotfish small (< 30 cm FL; TL 2.0-2.2); surgeonfish (*Acanthurus* spp.; TL 2.0-2.4); rabbitfish (*Siganus* spp.; TL 2.0-2.2); damselfish territories (farming species; TL 2.0 -- these are algae farmers that exclude other herbivores from territories and act as negative herbivores in some model implementations).

Invertivores: Wrasses large (*Cheilinus undulatus*, *Bodianus* spp.; TL 3.2-3.8); triggerfishes (Balistidae; TL 3.2-3.6); butterfly fishes (TL 2.8-3.2; coral polyp consumers). These groups are sensitive to the prey invertebrate community and serve as intermediate trophic level biomass indicators.

Zooplankton, phytoplankton, benthic primary producers (turf algae, macroalgae, CCA, zooxanthellae): The primary production compartments that determine the energy base of the food web. Macroalgae are the key output variable tracked in scenario simulations as the reef ecological state indicator.

#### STABLE ISOTOPE BASELINE VALUES (mean $\pm$ SD per region)

Indo-Pacific sites -- Turf algae baseline:  $\delta^{13}\text{C} = -18.4 \pm 1.4$  permil;  $\delta^{15}\text{N} = 3.4 \pm 0.8$  permil. CCA baseline:  $\delta^{13}\text{C} = -17.4 \pm 1.2$  permil;  $\delta^{15}\text{N} = 2.4 \pm 0.6$  permil. Zooplankton baseline:  $\delta^{13}\text{C} = -21.4 \pm 2.4$  permil;  $\delta^{15}\text{N} = 4.7 \pm 1.4$  permil.

Caribbean sites -- Turf algae baseline:  $\delta^{13}\text{C} = -17.4 \pm 1.8$  permil;  $\delta^{15}\text{N} = 2.8 \pm 0.8$  permil. *Diadema antillarum* (sea urchin; TL 2.0):  $\delta^{13}\text{C} = -14.7 \pm 1.4$  permil;  $\delta^{15}\text{N} = 6.4 \pm 1.4$  permil (consistent with benthic algae diet). Nassau grouper (*Epinephelus striatus*; apex piscivore):  $\delta^{15}\text{N} = 14.7 \pm 1.8$  permil (TL 4.1 from isotopes).

Red Sea sites -- Highest  $\delta^{15}\text{N}$  baseline (upwelling-influenced;  $\delta^{15}\text{N}$  baseline =  $5.4 \pm 1.4$  permil) requiring adjusted TP calculation:  $\text{TP} = 2 + (\delta^{15}\text{N}_{\text{consumer}} - 5.4) / 3.4$ . Reef sharks (*Carcharhinus melanopterus*) at Farasan:  $\delta^{15}\text{N} = 18.4 \pm 1.8$  permil (TL 4.6 from isotopes -- highest apex predator TL in dataset).